INTEGRATION OF MONOPOLAR AND BIPOLAR ELECTRODIALYSIS FOR VALORIZATION OF SEAWATER REVERSE OSMOSIS DESALINATION BRINES: PRODUCTION OF STRONG ACID AND BASE

4 M.Reig^{1,*}, S. Casas², C. Valderrama¹, O. Gibert^{1,2}, and J.L.Cortina^{1,2}.

⁵ ¹Chemical Engineering Dept. UPC-Barcelona TECH, Av. Diagonal 647 08028 Barcelona, Spain

6 ²CETAQUA Carretera d'Esplugues, 75, 08940 Cornellà de Llobregat, Spain

7 *Corresponding author: Tel.:+34 93 4016997; E-mail address: monica.reig@upc.edu

8

9 Abstract

10 Water scarcity in the Mediterranean basin has been solved by using seawater 11 desalination reverse osmosis technology (SWD-RO). This technology produces brine 12 which is discharged back into the sea resulting in an environmental impact on marine ecosystems. Under the circular economy approach, the aim of this work is to recover 13 14 resources from NaCl-rich brine (~60-70 g/L), e.g. in the form of NaOH and HCl, by integration of two ion exchange-based membrane technologies and quantify the 15 16 electrical energy consumption. Electrodialysis (ED) incorporating monovalent selective cation exchange membranes as divalent ions purification and concentration of the NaCl 17 present in the SWD-RO brine, was integrated with bipolar membrane ED (EDBM) to 18 produce NaOH and HCl. Current densities of 0.30-0.40 kA/m² at two temperature 19 20 ranges simulating different seawater temperature regimes (15-18 °C and 22-28°C) were tested and a pure NaCl solution was used as starting concentrate stream. NaCl-rich 21 22 brines with 100 or 200 gNaCl/L were obtained by ED and then introduced in the EDBM stack producing HCl and NaOH up to 2 M, depending on the initial concentrations. A 23 minimum energy consumption of 1.7 kWh/kgNaOH was calculated when working by 24 EDBM with initial concentrations of 104 g NaCl/L and 0.24 M HCl and NaOH. 25

26

27 Keywords: Desalination brines; Acid-base production; Electrodialysis; Bipolar28 membrane

29

1 **1. Introduction**

Water scarcity in the Mediterranean basin has been solved by using seawater and 2 reverse osmosis as desalination technology (SWD-RO). The process produces brine 3 which is discharged back into the sea resulting in an environmental impact on marine 4 ecosystems [1]. Initially, managing the generated brines was considered as an 5 6 environmental problem, but it has recently been identified as an opportunity to develop 7 circular economy concepts to transform a waste (the brine) to a resource for materials production [2]. Any medium to large SWD-RO plant drives so large amount of 8 dissolved elements that may be considered as a mine and then, the waterworks site 9 could be seen not only as a "water factory" but also as a potential chemicals production 10 site. Although revalorization of SWD-RO brines represents a promising and sustainable 11 12 alternative it has been applied so far only on a limited scale because of the large number 13 of technological gaps to be covered for making it economically feasible [3].

Seawater contains almost all elements in the periodic table [4]. However, only a few are 14 nowadays profitable extracted conventionally by evaporation: sodium chloride, 15 16 potassium chloride, magnesium and bromide salts. Several extraction schemes for a list of eight elements have been identified as being potentially economically and technically 17 18 viable (Na, K, Mg, Rb, P, Cs, In, Ge) [5]. Valorization approaches of SWD-RO brine to produce salts through different concentration/precipitation technologies have been 19 20 widely reviewed by Kim [6], Van der Bruggen et al. [7] and Pérez-González et al. [8] in an attempt to attain the zero liquid discharge (ZLD) objectives [9]. Due to the high 21 22 salinity and consequently high ionic conductivity, electrically driven membrane processes such as electrodialysis (ED) based on the selective passage of some 23 24 constituents through an ion-exchange membrane (IXM) have been widely researched. 25 Several studies have been oriented to the brine reuse by the chlor-alkali industry from SWD-RO brines [10–14]. This industry uses NaCl-rich brines to produce Cl₂, NaOH 26 and H₂ by means of ED as a salt valorization option. It was demonstrated that ED 27 technology concentrates NaCl from SWD-RO brines with competitive electrical 28 consumptions around 0.20 kwh/kg NaCl to produce 200 g NaCl/L. 29

Such solutions could be also used as raw materials to produce chemical commodities as
HCl and NaOH if bipolar membranes (BM) are integrated. However, ED combination
with bipolar membranes (EDBM) has been only applied successfully for chemical and

biochemical applications, food processing and in less extension for waste management 1 of metallurgical industries [1,15-19]. When electric current is applied between the 2 electrodes of the EDBM stack, water splitting is produced in the BM. Therefore, the 3 produced H^+ and OH^- ions can be used to generate acid (HX) and base (MOH) from 4 salts (MX), for example NaCl from the chemical industry to produce HCl and NaOH, 5 without production of hydrogen, oxygen or undesirable products. Another of the main 6 features of EDBM process is that water dissociation is accelerated up to 50 million 7 8 times compared to the rate of water dissociation in aqueous solutions. Moreover, 9 EDBM has low voltage drop, maximal energy utilization, space saving, easy installation and operation, low start-up and running costs and last but not least it can provide 10 products of high quality. However, one of the drawbacks of EDBM, when applied to 11 waste valorization, is that it is not as economically competitive as other membrane 12 13 separation technologies, due to the electrodes and ion exchange membranes cost and the capital cost. Despite this economical limitation, all the features previously 14 15 commented have made EDBM an environmentally friendly technology for valorization and management of industrial brines [1,20-22]. A clear application is, thus, the salt-rich 16 17 waste valorization for the production of acids and bases and a growth number of applications with industrial brines are reported [1,16,17,23,24]. However more limited 18 applications are devoted to SWD-RO brines to produce acid and base with EDBM at lab 19 or pilot scale [15,25]. These studies have concluded that the economic and technical 20 feasibility will be improved if the electrical consumption could be reduced with the 21 increase of brine concentration and with the reduction of scaling compounds in the 22 brines (HCO₃, Ca²⁺, Mg²⁺). In our previous study [26], SWD-RO brines (60 g/L NaCl) 23 were purified in divalent elements by using NF membranes and 1M HCl and 1 M 24 25 NaOH were produced by using EDBM with electrical consumptions of 2.6 kwh/kg NaOH. However, scarce data and studies could be found on the influence of the brine 26 27 concentration (e.g. NaCl) on the process efficiency in terms of the production of the highest NaOH and HCl concentrations and the specific electrical consumption. 28

The aim of this work is to quantify the performance and the electrical energy consumption of: a) the concentration and purification of NaCl from SWD-RO brines by using monovalent selective ion exchange membranes to avoid the presence of divalent metal ions (Ca, Mg) in ED cells; and b) the integration of an EDBM system for in-situ production of HCl and NaOH from concentrated NaCl brine (e.g. 100-200 g NaCl/L). The specific objectives are to find optimal operation conditions of the integrated
 processes of ED and EDBM to achieve the lowest specific electrical consumption and
 the highest acid and base concentrations.

4 2. Materials and methods

5 SWD-RO brine from the seawater desalination plant of El Prat (Barcelona, Spain) was 6 used as feed solution for the ED system. Mainly, this brine was rich in NaCl (65.1 ± 6.1 7 g NaCl/L), although it also contained other major components such as sulfate (5.4 ± 0.2 8 g/L), Mg²⁺ (2.6 ± 0.2 g/L), and Ca²⁺ (0.7 ± 0.04 g/L). Two different ranges of brine 9 temperature were evaluated: 15 to 18 °C from a spring season and 22 to 28 °C from a 10 summer season. Moreover, a pure NaCl solution was used as initial solution for the 11 concentrate loop.

In a previous study, Casas et al. [27] evaluated the removal of Ca^{2+} and Mg^{2+} as a 12 potential valorization pathway for seawater desalination brines. Treated brines 13 contained concentration of Ca^{2+} and Mg^{2+} below 10 mg/L. For practical purposes and 14 taking into account that the Ca-Mg purification process was evaluated previously, fresh 15 pure NaCl brines simulating the composition of the feed concentrated brines to be used 16 in the valorization as HCl and NaOH by EDBM were prepared in the present study. 17 18 Figure 1 shows also this procedure were pure NaCl was used as input brine for the ED concentrated loop. 19

Then, the ED cell was working with an open diluate loop and a NaCl recirculated concentrate stream. Sequentially, the NaCl-rich concentrated solution produced by the ED system was introduced in the EDBM cell which was used to produce NaOH and HCl. Details on the operation conditions are described as follows.



1

2 Figure 1. Schematic illustration of the evaluated seawater desalination brine valorization

3 process. Stream concentrations are indicated on the arrows.

4

5 2.1 Experimental description and operation of the ED plant.

An ED pilot using an Eurodia Aqualizer SV-10 stack (620x450x313 mm) with 50 cell
pairs made of Neosepta cation-exchange membranes (CIMS) and anion-exchange
membranes (ACS) (0.1 m² effective surface area per membrane) was used [13,14,28].
The main characteristics of the membrane are listed in Table 1.

10

11 Table 1. The main characteristics of the ED and EDBM membrane used.

	E	D	EDBM			
	ACS	CIMS	PC Acid 60	PC SK	PC BP	
Туре	Strongly basic anion permeable	Strongly acidic cation permeable	Strongly alkaline (ammonium)	Strongly acidic (sulfonic acid)	Water splitting	
General use	Monoanion permeselective (Cl-form)	Monocation permeselective (Na-form)	Monovalent acid (HCI/HNO ₃ /HF)	Standard desalination	efficiency > 95 %	
Electric	3.8	1.8	~ 2	~ 2.5	1	

resistance (Ω cm ²)					
Burst strength (MPa)	0.15	0.10	0.4 – 0.5	0.4 – 0.5	
Thickness (mm)	0.18	0.15	0.16 - 0.20	0.16 - 0.20	0.20 – 0.35

1

The intermembrane distance was 0.43 mm and linear flow velocity at the inlet of 2 3 desalting and concentrating cells was 11 cm/s. The feeding and the electrolyte circuits were operated in a single-pass design to achieve higher current densities and minimize 4 the problems of the increase of temperature in the cell. The concentrate (divalent-free 5 NaCl-rich) stream was re-circulated to reach the maximum NaCl concentration with 6 sustainable electrical specific consumption (e.g. current densities <0.4 kA/m² and 7 specific electric consumption <0.3kWh/KgNaCl) under the two given brine temperature 8 9 ranges evaluated. The SWD-RO brine flow rate through the stack was 0.5 m³/h in both the feeding and the concentrating stream compartments and 0.15 m^3/h in the electrodes 10 11 chambers. HCl was added to keep the pH below 4 for the cathodic circuit, below 7 in the feeding circuit and below 5.5 in the concentrate circuit. 12

Two temperature ranges simulating different seawater temperature regimes (15-18 °C and 22-28°C) were tested and current densities were varied between 0.3 and 0.40 kA/m² in order to obtain a NaCl-rich brine with the lower specific electrical consumption (less than 0.3 kWh/kg NaCl). The SWD-RO brine concentration process was monitored by in-line measurements of temperature, flow-rate, pressure, current intensity, voltage, electrical conductivity and pH as it can be seen in Figure 2.



Figure 2. ED pilot plant scheme. (P: pressure sensor, T: temperature sensor, C:
 conductivity sensor, F: flow-meter, pH: pH-meter, A: ammeter, V: voltmeter).

3

4 2.2 Experimental description an operation of the EDBM plant.

5 A lab-scale pilot incorporating an EDBM stack PCCell ED 64-004 (PCCell GmbH, Germany) was used. The dimension of the cell was 0.11x0.11m. A scheme of the 6 EDBM stack is shown in Figure 3. It was a 4 chamber system (electrode rinse, acid, 7 base and salt) with an active membrane area of 64 cm^2 per membrane. The stack 8 configuration was composed of three cell triplets; each cell triplet had one cationic 9 exchange membrane (CEM) (PC-SK), one anionic exchange membrane (AEM) (PC 10 11 Acid 60) and one bipolar membrane (BM). The main characteristics of the membrane are listed in Table 1. The EDBM cell worked under close loop configuration for the four 12 13 streams. Two electrodes rinse compartments formed a single circuit located at the cell ends. 14

Four pumps were used to impulse each stream into the EDBM unit. All the 15 performances were carried out at constant voltage (9 V) until the conductivity in the 16 feed tank was almost zero (values around 2 or 3 mS/cm). Some other parameters were 17 constant during the experiments, such as pressure drop differences between 18 compartments (0.3 bar) and the flow rates (ranged from 15 to 20 L/h for the acid, base 19 and salt stream and around 100 L/h in the electrode rinse stream). Several parameters, 20 such as, pressure, temperature, flow rate, electrical current, voltage and electrical 21 conductivity were monitored for all the circuits, while pH of the diluted solution was 22 23 also recorded.



Figure 3. EDBM pilot plant scheme. (P: pressure sensor, T: temperature sensor, C:
conductivity sensor, F: flow-meter, pH: pH-meter, A: ammeter, V: voltmeter, S.T: salt
tank, A.T: acid tank, B.T: base tank, E.T: electrode rinse tank).

Taking into account that high voltage and/or current density values implies higher acid and base final concentrations [1,15,25,26], the maximum stack voltage that could be applied (9 V) due to the membrane stack configuration was used. To apply this voltage threshold an initial acid and base solution had to be introduced in the corresponding tank.

10 Preliminary experiments were carried out with synthetic HCl, NaOH solutions. Also, NaCl solutions at two different concentration levels mimicking the concentrate solution 11 12 obtained as ED brines were used with a nominal concentration of 100 and 200 g NaCl/L 13 (1.71 and 3.42 M NaCl). Initial HCl and NaOH concentration were needed to avoid a high resistance in the power supply. HCl and NaOH feed solution for the acid and base 14 compartment respectively were also prepared with different concentrations (0.05, 0.10)15 and 0.50 M). For the electrode rinse compartment, a salt solution of nominal 16 concentration of 45 gNa₂SO₄/L (0.32 M Na₂SO₄) was prepared. 17

Different experiments were carried at a) different initial nominal NaCl concentration (100 g NaCl/L and 200 gNaCl/L) and b) different initial acid and base concentration (about 0.05, 0.10 and 0.50 M) to evaluate the effect of these parameters on the final concentration of HCl and NaOH produced. The experimental conditions used are
 collected in Table 2.

3

Table 2. Nominal experimental conditions including voltage, current density and initial
concentrations of the different EDBM stack circuits used.

Constant voltage	Current density	Electrode (Na₂SO₄)	Salt (NaCl)	Acid (HCl)	Base (NaOH)
(V)	(kA/m²)	C _{ini} E (g/L)	C _{ini} S (g/L)	C _{ini} AB (M)	C _{ini} AB (M)
			100 200	0.05	0.05
9 0.	0.5	45	100 200	0.10	0.10
			100 200	0.50	0.50

6

7 2.3 Analytical methodologies and chemical analysis

Samples from ED cell: samples from feed, diluate and concentrate streams were taken 8 every 2 hours. Chloride (Cl⁻) concentration was measured potentiometrically through 9 precipitation with AgNO₃ and an AgCl electrode by a Metrohm 721 instrument. Sulfate 10 $(SO_4^{2^-})$ concentration was measured by ionic chromatography using a Metrohm 761 11 Compact IC equipped with an Anion Dual 2–6.1006.100 column. Ca²⁺ and Mg²⁺ were 12 determined by atomic absorption spectrophotometry using a Perkin Elmer Analyst 300. 13 Finally, pH of all the samples was also measured using a glass electrode (Crison pH 14 Basic 20). 15

Samples from EDBM cell: samples from salt, acid, base and electrode rinse streams were taken before and during the experiments. HCl and NaOH samples were analyzed potentiometrically by acid-base titration (Titration Excellence T-70). Na⁺ cations and Cl⁻ and SO_4^{2-} anions concentration on the salt circuits were measured by ionic chromatography (Dionex ICS-1100 and ICS-1000, respectively).

21

2.4 Optimization of the energy consumption of the EDMB by response surface

2 methodology (RSM)

Response surface methodology (RSM) was used to determine the optimal initial salt, acid and base concentrations to obtain the minimum specific energy consumption by EDBM. A Central Composite Design (CCD) was used to determine the optimal conditions (e.g. production of concentrated NaOH and HCl solutions with the minimum energy consumption) for the critical factors (initial salt, acid and base concentrations). A response surface (Y) was built by fitting a second-order polynomial equation described by Equation 1:

$$Y = \beta + \sum_{i=1}^{2} \beta_i X_i + \sum_{i=1}^{2} \beta_{ii} X_i^2 + \sum_{i=1}^{2} \sum_{j=i+1}^{2} \beta_{ij} X_i X_j$$
(1)

where X_i and X_j represent the measured values of the independent variables, Y is measured response values or dependent variable and β_i s are model coefficients calculated from experimental data by using a fitting approach [29].

In this case study, the independent variables were the initial salt concentration ($C_{ini}S$) and the initial acid and base concentration ($C_{ini}AB$), while the response was the energy consumption ($Ec_{(EDBM)}$) in the EDBM stack. Therefore, the specific energy consumption ($Ec_{(EDBM)}$) describing the system could be described by Equation 2:

$$Ec_{(EDBM)} = \beta - \beta_1 \cdot C_{ini}AB + \beta_2 \cdot C_{ini}S + \beta_3 \cdot (C_{ini}AB)^2 + \beta_4$$

$$\cdot (C_{ini}S)^2 + \beta_5 \cdot C_{ini}AB \cdot C_{ini}S$$
(2)

where C_{ini}S is the initial salt concentration (gNaCl/L) and C_{ini}AB is the initial acid and
base concentration (M).

The range and levels of experimental variables (salt, acid and base concentrations) usedin this study are collected in Table 3.

- 21
- 22

23

24

1 Table 3. Coded levels and range of independent variables for experimental design

Level	Coded level	Fa	ctors
2070		C _{ini} S	C _{ini} AB
Low	-1	50*	0.05
Mid	0	100	0.1
High	1	200	0.5

* A 50 gNaCl/L solution was used in order to compare between three initial NaCl
concentrations levels.

4 2.5 Determination of energy consumption and current efficiency of NaCl 5 concentration in the ED process

The energy consumption (Ec_(ED)) for the NaCl concentration by ED was calculated as
the energy necessary to increase the NaCl concentration in the tank. It was calculated
according to Equation 3:

$$Ec_{(ED)}\left(\frac{kWh}{kg \ NaCl}\right) = \frac{U_{cell} \cdot I \cdot t}{v_{tank} \cdot C_{ini}S}$$
(3)

9

Where U_{cell} (V) is the mean membrane stack potential, I (A) is the applied current intensity, t (h) is the operation time, v_{tank} (L) is the NaCl solution volume of the concentrate tank and $C_{ini}S$ (gNaCl/L) is the sodium chloride concentration obtained at the end of the performance.

14

2.6 Determination of energy consumption and current efficiency in the EDBM process

Energy consumption (Ec_(EDBM)) and current efficiency (Ie) in the EDBM stack were
calculated by means of Equations 4 and 5, respectively.

$$Ec_{(EDBM)}\left(\frac{kWh}{kg \ product}\right) = \frac{(U \cdot I \cdot t)/1000}{m_{\text{final product}}}$$
(4)

Where U (V) and I (A) are the voltage and current applied in the EDBM stack, respectively, t (h) is the time of the operation and $m_{final product}$ (kg) is the acid or base produced.

$$Ie = \frac{F \cdot (V_t \cdot C_t - V_i \cdot C_i)}{I_d \cdot S \cdot t \cdot n}$$
(5)

Where F (96500 C/mol) is the faraday constant, V (L) and C (M) are the volume and concentration of the desired product at time "t" or at initial time "i", I_d (A/cm²) is the current density, S (64 cm²) the membrane surface, t (s) is the operational time and n is the cell triplets of the EDBM stack (3 in this study).

5

6 **3. Results and discussion**

7 **3.1** Concentration of NaCl brines by ED

The NaCl concentration profiles for different current densities (from 0.30 kA/m^2 to 0.48 kA/m²) at two different temperature ranges: a) 15 - 18 °C and b) 22 - 28 °C, are plotted 9 in Figure 4. Both temperature ranges provided the expected temperature regimes for 10 11 SWD-RO brines as consequence of the seasonal variation on the Mediterranean Sea. Current densities were limited to values below 0.4 kA/m² to reduce the specific 12 electrical consumption. For the two temperature scenarios the concentration of NaCl in 13 the concentrate tank increased gradually with time (from an initial 60-70 g NaCl/L) 14 until a plateau was reached by the end of the experiment at concentrations between 185 15 and 205 g NaCl/L for the low temperatures and between 105 to 135 g NaCl/L for the 16 high temperatures, depending on the current density applied: the higher the intensity 17 applied, the higher the final concentration of NaCl reached (Figure 4). 18



Figure 4. Evolution of NaCl concentration by ED. a) Experiments at 15 - 18 °C, b)
Experiments at 22 - 28 °C.

Firstly, the applied electrical potential favored the ion migration flux and the electro-5 osmosis flux. However, the NaCl concentration gradient between the diluate and the 6 concentrate circuits had an increase of 30 to 130 g/L NaCl, depending on the intensity 7 and temperature conditions. This NaCl concentration gradient promoted two new mass 8 transport phenomena: diffusion of NaCl from the concentrate compartment to the 9 10 diluate compartment and osmotic flux whereby water was transported from the diluate compartment to the concentrate one. These two mass transfer phenomena diminished 11 12 the desired effect of increasing the NaCl concentration on the concentrate tank and 13 increased the energy requirements of the process. Then, concentration experiments were 14 limited to the point where these phenomena were important. As seen in Figure 4, at the low temperature range (15 - 18 °C), NaCl concentrations between 185 and 205 gNaCl/L 15

were achieved after 12 hours of operation, whereas only between 105 and 135 gNaCl/L
at the high temperature range (22 - 28 °C). Lengthening the experiment time did not
provide any substantial increase in the NaCl concentration (increase differences less
than approximately 1 g/L) after 12 hours of additional concentration.

Experiments at both temperature regimes have shown the temperature influence on ED 5 6 mass transport processes and properties of the brine (electrical conductivity and 7 density), solvent (molecular diffusivity) and membrane (resistance and electrical conductivity) [30,31]. The temperature influence on the concentration process in ED is 8 more complex than that exerted by the current intensity because the different mass 9 transfer processes involved (e.g., it may be because the osmosis transport of NaCl) are 10 temperature dependent. Although theoretical models to describe this dependence are 11 12 still lacking, Tanaka et al. [32] have recently developed a model for the concentration of seawater to produce salt. This model was based on analysis of a large data base 13 14 containing experimental results of ion transport through ion exchange membranes where the correction of temperature for physical parameters, such as density, electrical 15 16 conductivity and membrane resistance was considered.

Experiments also showed that it was possible to produce a NaCl-rich concentrate stream 17 from SWD-RO brine by means of an ED pilot plant. Due to the single-pass design 18 circuit for the diluate stream, and the recirculated NaCl-rich concentrate loop, a final 19 divalent free solution containing concentrated NaCl was obtained. Moreover, as it was 20 reported previously [14], the rejection of divalent minor ions, such as Ca^{2+} , Mg^{2+} and 21 SO₄²⁻, was favored in the concentrate stream due to the selectivity of the ion exchange 22 membranes used. In this study, final concentration of Ca^{2+} and Mg^{2+} were 0.31 and 0.30 23 g/L, respectively for the high temperature experiments and 0.55 g Ca^{2+}/L and 0.49 g 24 Mg^{2+}/L for the low temperature experiments. Then, it can be stated that the divalent ions 25 migration flux was insignificant in the concentrate loop. Yang et al. [15] also used a 26 pre-treatment of the seawater RO concentrate to reduce Ca²⁺ and Mg²⁺ ions and avoid 27 scaling using Na₂CO₃ and NaOH to precipitate Ca^{2+} as CaCO₃ and Mg²⁺ as Mg(OH)₂. 28 After precipitation, the pH of the RO concentrate was adjusted with HCl to prevent 29 organic and inorganic fouling on the anion exchange membranes of the EDBM stack. 30

Recently, the integration of nanofiltration (NF) and EDBM for the valorization of SWD-RO brine has been evaluated [26], where NF was applied for the removal of

divalent ions from the SWD-RO brine. NF rejected 20% of NaCl, resulting in an undesired dilution of the NaCl in the treated brine. In the present study, the application of ED allowed not only to avoid the presence of divalent ions in the treated brine but also to increase the nominal NaCl concentration from 65 g/L up to 100 and 200 g/L, depending on the operation conditions. With this increase of NaCl concentration, it was possible to achieve higher HCl and NaOH concentrations when compared with those obtained using NF as purification step.

8

9 **3.2 Energy consumption on the concentration of NaCl (Ec_(ED))**

The energy consumption $(Ec_{(ED)})$ raised gradually in all the experiments until the 10 concentration in the tank reached its maximum concentration depending on the 11 operation conditions (temperature and current density). Experiments were stopped when 12 13 the energy consumption increased dramatically because osmosis and ion diffusion fluxes inside the stack reached a maximum due to the concentration gradient between 14 the diluate and the concentrate compartments (data not shown). A summary of the 15 energy consumption for continuous operation ($Ec_{(ED)}$) for different operation conditions 16 evaluated is shown in Table 4. 17

18

Table 4. NaCl concentration achieves and energy consumption calculated depending onexperimental conditions

Temperature (ºC)	Current density (kA/m²)	NaCl concentration, C _{ini} S (g/L)	Energy consumption Ec _(ED) (kWh/kg NaCl)
16.6	0.40	204.8	0.217
17.9	0.35	185.4	0.154
18.2	0.30	187.7	0.148
21.7	0.40	135.1	0.096
22.0	0.40	133.9	0.055
28.3	0.30	104.0	0.161

21

The energy consumption increased if working with higher current density, thus, 0.4kA/m² was confirmed as the maximum operable point. The temperature affects the product flow rate, and lower flow rates were obtained at higher temperatures.

An average of 0.17 kWh/kg NaCl was calculated when around 200 gNaCl/L were 1 produced, while lower values around 0.10 kWh/kg NaCl were achieved when only 2 approximately 100 gNaCl/L were concentrated by ED. The values obtained in our pilot 3 plant were within the range of the published results. For laboratory scale studies [33], 4 the energy consumption with NaCl solutions up to approximately 100 g/L ranged 5 between 0.18 and 0.33 kWh/kg with 80% solute recovery and current densities of 0.20 6 to 0.30 kA/m² were reported. Besides, it was reported that an ED industrial target in 7 Japan was to produce 200 g NaCl/L from seawater with an energy consumption lower 8 9 than 0.12 kWh/kg NaCl [34].

10 It can be concluded that concentration of NaCl brines by ED is dependent on the temperature and current density applied. Higher current density and low temperature 11 12 lead to higher NaCl concentration. For this reason, in order to achieve 200 gNaCl/L by ED with specific electrical consumption below 0.2 kWh/kg NaCl, it was preferable to 13 14 work during winter-spring seasons (when temperatures are below 20°C) and at high current densities. However, during summer season (when temperatures are above 21°C) 15 NaCl brines between 100 to 130 g/L were obtained at specific electrical consumption 16 around 0.1 kWh/kg when using current densities of 0.30 and 0.40 kA/m². 17

18

3.3 Evaluation of HCl and NaOH production from NaCl brines by means of EDBM

EDBM (Figure 5) experiments using NaCl brines (nominal concentration of 100-200 21 gNaCl/L) as feed solution indicated that the desalinating process and the production of 22 NaOH and HCl occured efficiently as was previously reported [16,17]. For both brines 23 100 and 200 gNaCl/L, NaCl concentrations at the end of the experiments were around 24 25 20 mgNaCl/L. It is important to point out that conductivity of the electrode rinse stream 26 (Na₂SO₄) was constant along the whole experiment. This fact indicated that no leak or non-desired ion transport occurred between the electrode rinse compartment and their 27 28 adjacent compartments. Figure 5 also shows the final NaOH and HCl concentrations obtained with an initial nominal NaCl concentration of 100 g/L (1.71 M NaCl) (Figure 29 30 5a) and 200 g/L (3.42 M NaCl) (Figure 5b) as a function of the initial concentrations evaluated for the acid and base circuits (from 0.05 to 0.5 M). 31



Figure 5. Salt, acid and base concentration at the initial and final experimental time
under different initial concentration of HCl and NaOH (0.05, 0.10 and 0.50 M). Using
initial nominal NaCl concentration of a) 100 g/L and b) 200 g/L.

Figure 5 demonstrates that the initial acid and base concentration had no significant
effect on their final concentration, in contrast with initial NaCl concentration, which
appeared to effect the final concentration of NaOH and HCl. In fact, working with
initial NaCl 100 g/L (1.71 M) allowed to concentrate NaOH and HCl concentrations up
to 1.26 M and 1.65 M, respectively, whereas using the concentrate stream of ED of 200
g NaCl/L (3.42 M) allowed to achieve final concentrations of NaOH and HCl around 2
M.

The NaCl percentage that was converted into NaOH and HCl for the different
 experiments is shown in Table 5. When current is applied between the electrodes,
 proton ions have higher mobility than hydroxyl ones, then, conversion percentage to
 HCl could be higher than the NaCl percentage that was converted into NaOH.

5

Table 5. NaCl percentage that was converted into NaOH and HCl in the first set ofEDBM experiments.

Initial HCl and NaOH	Initial salt co 100 g l	ncentration: NaCl/L	Initial salt concentration: 200 g NaCl/L		
concentration	NaOH	HCI	NaOH	HCI	
0.05	65 %	84 %	57 %	61 %	
0.10	69 %	75 %	64 %	59 %	
0.50	46 %	68 %	52 %	54 %	

As it can be seen in Table 5, the higher conversion percentage was achieved working
with the nominal 100 g NaCl/L solution and 0.05 M HCl and NaOH. In general, higher
percentages were obtained with 100 gNaCl/L solutions and higher conversion ratios to
HCl than to NaOH reaching a maximum of 84 %.

The influence of the initial NaCl concentration (~100 and 200 gNaCl/L) on the final acid and base concentrations for 0.05M HCl and NaOH initial solutions is shown in Figure 6. In order to make a better comparison, some experiments using 50 g NaCl/L (approaching the SWD-RO brine composition) as feed solution were conducted. Then, the results are also shown in Figure 6 comparing the initial and final concentrations obtained.



1 2

Figure 6. Salt, acid and base concentration at the initial and final experimental time
under different initial concentration of NaCl

Figure 6 shows that working with an initial NaCl concentration of 50 g/L it was possible to obtain NaOH 0.66 M and HCl 0.90M. These results are in agreement with those reported by Ibañez et al. [25], who obtained very similar HCl and NaOH concentrations (0.71 and 0.91M, respectively) under the same current density (0.5 kA/m²) and initial NaCl concentration of 50 g/L. The higher concentration values obtained for HCl in comparison with those for NaOH were associated to the higher exchange capacity of the anionic membranes and the lower co-ion transport through them.

It should be pointed out that it was possible to raise the final HCl and NaOH 11 12 concentration by increasing the initial NaCl concentration. As it is shown in figure 6, by using a more concentrated NaCl feed (~100 g/L) the NaOH and HCl concentrations 13 14 obtained were 1.2 and 1.5 M, respectively, and for NaCl solutions of 200 g/L, it was possible to concentrate NaOH up to 1.9 M and HCl up to 2.0 M. Thus, higher HCl and 15 16 NaOH concentrations can be obtained by working at high current density and high initial NaCl concentration. As described previously, Yang et al. [15] used a pretreated 17 18 seawater RO concentrate as feed solution using a similar EDBM configuration and reached 0.4 M NaOH, whereas Ibañez et al. [25] obtained NaOH up to 1 M working 19 20 with synthetic SWD-RO concentrate.

The maximum NaOH concentration obtained in this work was 2 M (8% w/v), which agrees with previous published studies where the maximum NaOH concentration reached by EDBM at bench- and pilot-scale ranged from 8 to 9.5% (and up to 12.5% if
the electrochemical consumption was increased) [23–25]. These values reported by
bench- and pilot-scale systems are, however, low compared to the one reached by the
current chlor-alkali industry (32-50 % w/v depending on the technology applied).

5 The initial approach is the on-site use of the NaOH produced or the use by other 6 industries with distances below 40 km (e.g., reagent for pH adjustment or other 7 commodities uses). If the objective is out-side use, the production site should consider a 8 concentration step up to 32-50% by evaporation with an electrical consumption of 9 2400-4500 kWh/t NaOH [35].

3.4 Evaluation of the energy consumption on the HCl and NaOH production from NaCl brines by means of EDBM

Table 6 summarizes the results of the different assays performed on the production of HCl and NaOH from NaCl brines by the EDBM stack including the initial and final concentrations of each stream and also the quantity of HCl produced and the energy consumption to produce NaOH under a voltage of 9 V. Although HCl and NaOH were obtained from NaCl, taking into account the lowest market prices of HCl in comparison to NaOH, energy consumption ($Ec_{(EDBM)}$) was calculated as kWh per kilogram of NaOH obtained, and HCl was considered a by-product.

19 Table 6. Experimental results on the specific energy consumption on the production of

20 NaOH and HCl from NaCl solutions (~ 50, 100 and 200 gNaCl/L) by EDBM at room

21 temperature (21 ± 1 °C).

	Ĩ	Electr (Na₂S	ode O ₄)	Sa (Na	lt Cl)	Ac (H	cid Cl)	Base (NaOH)		Time		Ec _(EDBM)
(V) (kA/ (V) m ²)	(kA/ m²)	C _{initial} (g/L)	C _{final} (g/L)	C _{initial} (g/L)	C _{final} (g/L)	C _{initial} (M)	C _{final} (M)	C _{initial} (M)	C _{final} (M)	(h)	kg HCl produced	kWh/kg NaOH produced
9.0	0.41	45.50	43.59	53.60	0.01	0.06	0.90	0.05	0.66	2.44	0.031	2.345
8.9	0.47	42.46	43.37	99.58	0.03	0.07	1.49	0.05	1.16	4.07	0.052	2.460
8.9	0.52	42.37	47.95	186.15	0.02	0.07	2.01	0.05	1.87	6.75	0.071	2.713
8.9	0.44	44.66	43.85	52.98	0.01	0.12	0.88	0.10	0.71	2.54	0.028	2.576
8.8	0.41	43.25	41.93	98.18	0.02	0.11	1.35	0.09	1.25	3.63	0.045	1.822
8.9	0.52	47.48	45.92	186.88	0.02	0.10	1.99	0.10	2.14	6.24	0.069	2.238
8.9	0.47	46.86	45.03	52.88	0.02	0.53	1.16	0.44	1.06	2.68	0.023	2.888
8.9	0.43	44.86	49.19	97.21	0.02	0.54	1.65	0.50	1.26	3.89	0.041	3.104
8.9	0.59	41.45	45.21	186.61	0.02	0.55	2.23	0.41	2.03	6.94	0.061	3.624

As it reported in the last column of Table 6, a range of $Ec_{(EDBM)}$ values were achieved from 1.822 up to 3.624 kWh/kg NaOH. In order to define the initial acid, base and salt concentrations that require the minimum $Ec_{(EDBM)}$, a factorial design was defined to obtain a general lineal model using the response surface methodology (RSM) and central composite design (CCD). Table 7 represents the least squares matrix (Z) with different values of initial salt, acid and base concentration and the response matrix (Y) of energy consumption.

			Z			Y
	C _{ini} AB	C _{ini} S	C _{ini} AB ²	C _{ini} S ²	$\textbf{C}_{ini}\textbf{AB} {\boldsymbol{\cdot}} \ C_{ini}S$	Ec _(EDBM)
1	0.05	50	0.0025	2500	2.5	2.34
1	0.05	100	0.0025	10000	5	2.46
1	0.05	200	0.0025	40000	10	2.71
1	0.1	50	0.01	2500	5	2.58
1	0.1	100	0.01	10000	10	1.82
1	0.1	200	0.01	40000	20	2.24
1	0.5	50	0.25	2500	25	2.89
1	0.5	100	0.25	10000	50	3.10
1	0.5	200	0.25	40000	100	3.62

8 Table 7. Initial matrix for the statistic model (Z) and response matrix (Y)

9

The model coefficients (β) in the factorial design, described by the matrix in Table 7,
were estimated by Equation 6: .

$$\beta = [(Z \cdot Z')^{-1}] \cdot [Z' \cdot Y] \tag{6}$$

Then, the initial salt and acid/base concentration values to obtain the minimum energyconsumption could be estimated by Equation 7:

$$Ec_{(EDBM)} = 3.4589 - 9.7568 \cdot C_{ini}AB - 0.0116 \cdot C_{ini}S + 18.5658 \cdot (C_{ini}AB)^{2} + 4.5173e - 5 \cdot (C_{ini}S)^{2} + 0.0094 \cdot C_{ini}AB \cdot C_{ini}S$$
(7)

14

A surface plot for $Ec_{(EDBM)}$ values as a function of salt and acid/base concentrations, including both the experimental (by un-filled circles) and model predicted values, is shown in Figure 7. Figure 7 shows that for the used EDBM stack the optimum operating values yielding a minimum energy consumption of 1.7 kWh/kg NaOH, at a constant voltage of $8.9\pm0.1V$, can be obtained with and initial NaCl concentration of 104 ± 4 g/L and initial acid and base concentration of 0.24 ± 0.2 M.



1

Figure 7. Surface plot for optimal values of energy consumption versus initial salt andacid/base concentration.

Scarce data of process performance and energy consumption for application of BM 4 stacks with brines can be found in the scientific literature. Table 8 summarizes the 5 published data on brine treatment by means of EDBM for acid and base production. 6 Different brine types had been used as feed solution with different concentrations. Most 7 of them had been pre-treated in order to be introduced in the EDBM stack by means of 8 different techniques. Also, the EDBM cell used included different dimensions, active 9 10 surface membrane area and membrane properties (e.g. different manufacturers). Some 11 authors had calculated the energy consumption of the EDBM process. It is important to mention that it has not been found any published work on the integration of a brine pre-12 13 concentration step before the EDBM, in which the total energy consumption was calculated. Then, in this work apart from the reported energy consumption of the 14 production of NaOH and HCl by the EDBM process, the energy consumption of the ED 15 pre-treatment has also been taken into account: around 0.17 kWh/kg NaCl when the 16 brine was concentrated up to 200 gNaCl/L and 0.10 kWh/kg NaCl when concentrated 17 up to 100 g NaCl/L. 18

Brine Type	Brine Composition (g/L)	EDBM cell (Area (m²))	By-products	Current efficiency (%) / Current density (kA/m ²)	Energy consumption	Ref.
SWD-RO brine.	Cl (38.8), Na (20.8), SO ₄	PCA-SK (0.0064)	HCI 1.99 M	55-88 / 0.5	1.8-3.6 kWh/kg NaOH	This
Concentrated by ED	(5.41), Mg (2.64), Ca (0.83).	PCA-Acid 60	NaOH 2.14 M		1.9 - 3.8 kWh/kg HCl	work
	EDBM feed: NaCl (100;200)	PCA-BM	(200gNaCl/L; 0.5 kA/m ²)		0.08 - 0.23 kWh	
Synthetic SWD-RO	NaCl (54), K (0.7), Mg (-),	RALEX AMH (0.02)	HCI 0.6-0.8 M	80 / 0.25	n.a	[25]
concentrate	Ca (-), SO ₄ (5.3), HCO ₃ (0.2)	RALEX CMH	NaOH 0.7-1.0 M	50 / 1.0		
		NEOSEPTA-BP1	(0.25-1 KA/m ²)			
SWD-RO/ Pre-	NaCl (38), K (0.4), Mg (-),	Qianqiu AEM (0.0088)	HCI 0.7 M	50-66 / 0.35-	7.2-9.4 kWh/kg HCl	[15]
treated for Ca/Mg	Ca (-), SO ₄ (3.8),	Qianqiu CEM	(0.57 kA/m ²)	0.57		
removal	HCO ₃ (0.07), Br(0.1)	Fumatech – BPM	NaOH 0.4 M			
Sea water	n.a.	Commercial : (0.012)	Commercial:	Commercial:	Commercial:	[36]
		CMI - 7000S	HCI 0.4 M	15-45 / 0.05	0.129 kWh	
		AMI - 7001S	NaOH 0.5 M			
		PSDVB – BPM	Synthetic:	Synthetic:	Synthetic:	
		Synthetic: (0.012)	HCI 0.6 M	35-63 / 0.05	0.14 kWh	
		PSu - CEM, AEM, BPM	NaOH 0.7 M			
Industrial Brine	NaCl (70), Ca (-), TIC (0.01),	PCA-SK (0.0064)	PBM: (0.572 kA/m ²)	PBM: 65 / 0.572	n.a.	[23]
	TOC (0.132), pH=12.6	PCA-Acid 60	HCl 1.6 M, NaOH 1.7 M	FBM: 81 / 0.523		
		Fumatech – FBM	FBM: (0.523 kA/m ²)			
		PCA-PBM	HCl 1.8 M, NaOH 2.0 M			
Industrial Brine	NaCl (20),SO4 (2.9), K (6.6),	PCA-SK (0.0064)	HCI 1.6 M	Cl: 64 / 0.6	n.a.	[17]
	Ca (0.8), TIC (0.01), TOC	PCA-Acid 60	H ₂ SO ₄ 0.12 M	SO ₄ : 10 / 0.6		
	(0.02), pH=7.6.	PCA-BM	NaOH 1.2 M	Na: 47 / 0.6		
	Ca removal pre-treatment		KOH 0.7 M	K: 31 / 0.6		
Industrial Brine	SO4(89), Ca(0.02), Na(45)	Fumatech FAB	0.29 – 0.59 M mix acid	Mix acid:	5 – 6 kWh/kg mix acid	[24]
	Ni (0.01), pH=10.5	(0.0064)	0.34 – 0.68 M mix base	60 / 0.3; 70 / 0.6	4 -5 kWh/kg mix base	
	Pre-treatment: pellet	Fumatech FKB	(0.3 to 0.6 kA/m ²)	Mix base:		
	reactor	Fumatech – FBM		80 / 0.3; 90 / 0.6		

1 uole 0. Comparison of the published data on office reachent by LDDM

As it can be seen in Table 8, the highest HCl and NaOH concentration was achieved in 1 this work. This was possible due to the pre-treatment concentration by ED process. The 2 initial feed for the EDBM in this case was the higher among the reviewed studies. For 3 the current efficiency, it can be stated that similar values (e.g. from 60-88%) were 4 calculated in comparison to the other studies. As discussed previously on the literature, 5 the observed trend is that the amount of electric current converted into the desired 6 7 product (current efficiency) decreases with increasing current density within the evaluated interval working with synthetic SWD-RO concentrate [25], synthetic sea 8 9 water brines [37] or synthetic brackish brines [1]. High values of current density increased the concentration of products more rapidly by leading to an increase of a co-10 11 ion leakage. Finally, regarding the energy consumption estimation, each study took into account different parameters to be referenced. For example, Venugopal et al. [36] 12 13 reported as total kWh consumed, and values calculated were in the same range as the obtained in the present study. Besides, Yang et al. [15] stated it as kWh per each kg of 14 HCl produced. Their energy consumption values were higher (7.2 - 9.4 kWh/kg HCl) 15 than the ones calculated in this work (1.9 - 3.6 kWh/kg HCl). Also, Tran et al. [24] 16 17 quantified as kWh per each kg of mix acid obtained, but also as kWh per kg of mix base produced. In both cases, the reported energy consumption (5 -6 kWh/kg mix acid and 4 18 -5 kg mix base) were higher than in this study (1.9 - 3.6 kWh/kg HCl and 1.8 - 3.6 19 kWh/kg NaOH). Only, recently, Wang et al. [38], developed a mathematical model of a 20 typical three-compartment EDBM process to calculate the energy consumption and total 21 cost of the process. With an increase in the current density, the energy consumption of 22 the process increases and the total process cost decreases at first and then increases 23 gradually with a minimum cost value at a given current intensity. 24

25

26 **4.** Conclusions

ED was used in order to pre-concentrate SWD-RO brines from 60-70 g/L up to 100 or
200 g NaCl/L by appropriate current density and temperature control. Low temperatures
helped to concentrate easily the SWD-RO brines up to 200 g NaCl/L, while lower
concentrations (100 g NaCl/L) were achieved at higher temperatures (more than 20 °C).
The ED energy consumption for 0.30 – 0.40 kA/m² current densities ranged from 0.10
to 0.17 kWh/kg NaCl produced, depending on the final NaCl concentration achieved.

Moreover, by working with a NaCl initial concentrated stream it was possible to obtain
 a NaCl-rich solution with reduced levels of divalent ions.

3 According to the EDBM results using NaCl concentrates, HCl and NaOH solutions up to 2 M concentration were produced. EDBM technique requires an initial acid, base and 4 salt concentration. The initial acid and base concentrations represented no substantial 5 effect on the overall performance. Then, the minimum initial HCl and NaOH 6 7 concentrations are advised to be used. Moreover, concerning the effect of feed salt concentration, it was reported that an increase of NaCl led to an increase of acid and 8 base concentration. Then, the ED process can be a successful pre-treatment in order to 9 have a more concentrated initial feed in the EDBM and produce a more concentrated 10 HCl and NaOH. 11

Energy consumption measured in EDBM was ranged from 1.8 to 3.6 kWh/kg NaOH. 12 Consumption values were lower to those reported in the literature for similar brines 13 (e.g., pre-treated RO concentrates). Finally, the lineal model developed identified a 14 minimum EDBM energy consumption of 1.7±0.1 kWh/kg NaOH for an initial NaCl 15 16 concentration of 104±4 gNaCl/L and for the initial concentration of HCl and NaOH circuits of 0.24±0.04 M. Although the energy consumption values for the production of 17 acid and base to valorize brines are above the market prices, only for in-situ waterworks 18 uses or when industries would be forced to follow more sustainable brine management 19 20 approaches the proposed solution would have potential interest for implementation.

21

22 Acknowledgments

This research received support by the ZERO-DISCHARGE project (CTQ2011-26799) 23 24 and the Waste2Product (CTM2014-57302-R) financed by the "Ministerio de Economía y Competitividad" and the Catalan Government (Project Ref. SGR2014-50-SETRI), 25 Spain. The work of Monica Reig was supported by the Spanish Ministry (MINECO) 26 within the scope of the grant BES-2012-051914. We want to thank the contribution of 27 S. Asensio and C. Aladjem (Solvay Ibérica) and Dr. Fernando Valero (Aigües Ter 28 Llobregat) for the discussion on the results. We would also like to acknowledge M. 29 Martinez for her contribution to the project. 30

5. References

2 3 4	[1]	M. Badruzzaman, J. Oppenheimer, S. Adham, M. Kumar, Innovative beneficial reuse of reverse osmosis concentrate using bipolar membrane electrodialysis and electrochlorination processes, J. Memb. Sci. 326 (2009) 392–399.
5 6	[2]	European Comission, Moving towards a circular economy, Circ. Econ. Strateg. Roadmap. (2015) 1–9.
7 8 9	[3]	S. Adham, J. Oppenheimer, M. Kumar, Innovative Approaches to RO Concentrate Management: Beneficial Reuse and Concentrate Minimization, Water Environ. Fed. (2006) 4334–4344.
10 11 12	[4]	H. Ohya, T. Suzuki, S. Nakao, Integrated system for complete usage of components in seawater: A proposal of inorganic chemical combinat in seawater, Desalination. 134 (2001) 29–36.
13 14	[5]	T. Jeppesen, L. Shu, G. Keir, V. Jegatheesan, Metal recovery from reverse osmosis concentrate, J. Clean. Prod. 17 (2009) 703–707.
15 16	[6]	D.H. Kim, A review of desalting process techniques and economic analysis of the recovery of salts from retentates, Desalination. 270 (2011) 1–8.
17 18 19	[7]	B. Van Der Bruggen, L. Lejon, C. Vandecasteele, Reuse, treatment, and discharge of the concentrate of pressure-driven membrane processes, Environ. Sci. Technol. 37 (2003) 3733–3738.
20 21 22	[8]	A. Pérez-González, A.M. Urtiaga, R. Ibáñez, I. Ortiz, State of the art and review on the treatment technologies of water reverse osmosis concentrates, Water Res. 46 (2012) 267–283.
23 24 25 26	[9]	US Department of the Interior Bureau of Reclamation, Evaluation and Selection of Available Processes for a Zero-Liquid Discharge System for the Perris , California , Ground Water Basin, Desalin. Water Purif. Res. Dev. Progr. Rep. No. 149. (2008).
27 28 29	[10]	Y. Zhang, K. Ghyselbrecht, B. Meesschaert, L. Pinoy, B. Van der Bruggen, Electrodialysis on RO concentrate to improve water recovery in wastewater reclamation, J. Memb. Sci. 378 (2011) 101–110.
30 31	[11]	E. Korngold, L. Aronov, N. Daltrophe, Electrodialysis of brine solutions discharged from an RO plant, Desalination. 242 (2009) 215–227.

1 2 3 4	[12]	S. Casas, C. Aladjem, J.L. Cortina, E. Larrotcha, L. V Cremades, Seawater reverse osmosis brines as a new salt source for the chlor-alkali industry: integration of NaCl concentration by electrodialysis, Solvent Extr. Ion Exch. 30 (2012) 322–332.
5 6 7 8	[13]	S. Casas, N. Bonet, C. Aladjem, J.L. Cortina, E. Larrotcha, L. V Cremades, Modelling sodium chloride concentration from seawater reverse osmosis brine by electrodialysis: Preliminary results, Solvent Extr. Ion Exch. 29 (2011) 488– 508.
9 10 11	[14]	M. Reig, S. Casas, C. Aladjem, C. Valderrama, O. Gibert, F. Valero, et al., Concentration of NaCl from seawater reverse osmosis brines for the chlor-alkali industry by electrodialysis, Desalination. 342 (2014) 107–117.
12 13 14	[15]	Y. Yang, X. Gao, A. Fan, L. Fu, C. Gao, An innovative beneficial reuse of seawater concentrate using bipolar membrane electrodialysis, J. Memb. Sci. 449 (2014) 119–126.
15 16	[16]	S. Koter, A. Warszawski, A new model for characterization of bipolar membrane electrodialysis of brine, Desalination. 198 (2006) 111–123.
17 18 19 20	[17]	K. Ghyselbrecht, M. Huygebaert, B. Van der Bruggen, R. Ballet, B. Meesschaert, L. Pinoy, Desalination of an industrial saline water with conventional and bipolar membrane electrodialysis, Desalination. 318 (2013) 9– 18.
21 22	[18]	R. Ibáñez, P. Mier, M.J. Rivero, I. Ortiz, Electrodiálisis con membranas bipolares. Fundamentos y aplicaciones., Ing. Química. 418 (2004) 166–182.
23 24	[19]	M. Fidaleo, M. Moresi, Electrodialysis applications in the food industry., Adv. Food Nutr. Res. 51 (2006) 265–360.
25 26	[20]	T. Xu, C. Huang, Electrodialysis-based separation technologies: A critical review, AIChE J. 54 (2008) 3147–3159.
27 28 29	[21]	P. Saremirad, H.G. Gomaa, J. Zhu, Effect of flow oscillations on mass transfer in electrodialysis with bipolar membrane, J. Memb. Sci. 405-406 (2012) 158–166.
30 31	[22]	H. Strathmann, Electrodialysis, a mature technology with a multitude of new applications, Desalination. 264 (2010) 268–288.

1 2 3	[23]	K. Ghyselbrecht, A. Silva, B. Van der Bruggen, K. Boussu, B. Meesschaert, L. Pinoy, Desalination feasibility study of an industrial NaCl stream by bipolar membrane electrodialysis, J. Environ. Manage. 140 (2014) 69–75.
4 5 6 7	[24]	A.T.K. Tran, P. Mondal, J. Lin, B. Meesschaert, L. Pinoy, B. Van der Bruggen, Simultaneous regeneration of inorganic acid and base from a metal washing step wastewater by bipolar membrane electrodialysis after pretreatment by crystallization in a fluidized pellet reactor, J. Memb. Sci. 473 (2015) 118–127.
8 9 10	[25]	R. Ibáñez, A. Pérez-González, P. Gómez, A.M. Urtiaga, I. Ortiz, Acid and base recovery from softened reverse osmosis (RO) brines. Experimental assessment using model concentrates, Desalination. 309 (2013) 165–170.
11 12 13 14	[26]	M. Reig, S. Casas, O. Gibert, C. Valderrama, J.L. Cortina, Integration of nanofiltration and bipolar electrodialysis for valorization of seawater desalination brines: Production of drinking and waste water treatment chemicals, Desalination. 382 (2016) 13–20.
15 16 17 18	[27]	S. Casas, C. Aladjem, E. Larrotcha, O. Gibert, C. Valderrama, J.L. Cortina, Valorisation of Ca and Mg by-products from mining and seawater desalination brines for water treatment applications, J. Chem. Technol. Biotechnol. 89 (2014) 872–883.
19 20 21	[28]	S. Casas, C. Aladjem, J.L. Cortina, E. Larrotcha, L. V Cremades, Seawater RO Brine Reuse in the Chlor-alkali industry: Preliminary Results on NaCl Concentration By Electrodialysis, Solvent Extr. Ion Exch. 30 (2012) 322–332.
22 23 24 25	[29]	S. Ghasempur, SF. Torabi, SO. Ranaei-Siadat, M. Jalali-Heravi, N. Ghaemi, K. Khajeh, Optimization of Peroxidase-Catalyzed Oxidative Coupling Process for Phenol Removal from Wastewater Using Response Surface Methodology, Environ. Sci. Technol. 41 (2007) 7073–7079.
26 27 28	[30]	Y. Tanaka, Ion-exchange membrane electrodialysis for saline water desalination and its application to seawater concentration, Ind. Eng. Chem. Res. 50 (2011) 7494–7503.
29 30	[31]	Y. Tanaka, Ion-exchange membrane electrodialysis of saline water and its numerical analysis, Ind. Eng. Chem. Res. 50 (2011) 10765–10777.
31 32	[32]	Y. Tanaka, M. Reig, S. Casas, C. Aladjem, J.L. Cortina, Computer simulation of ion-exchange membrane electrodialysis for salt concentration and reduction of

1		RO discharged brine for salt production and marine environment conservation,
2		Desalination. 367 (2015) 76–89.
3	[33]	M. Fidaleo, M. Moresi, Optimal strategy to model the electrodialytic recovery of
4		a strong electrolyte, J. Memb. Sci. 260 (2005) 90–111.
5	[34]	Y. Tanaka, A computer simulation of feed and bleed ion exchange membrane
6		electrodialysis for desalination of saline water, Desalination. 254 (2010) 99–107.
7	[35]	T. Brinkmann, G. Giner santonja, F. Schorcht, S. Roudier, L. Delgado sancho,
8		Best Available Techniques (BAT) reference document for the production of
9		chlor-alkali, Luxembourg, 2014.
10	[36]	K. Venugopal, S. Dharmalingam, Desalination efficiency of a novel bipolar
11		membrane based on functionalized polysulfone, Desalination. 296 (2012) 37-45.
12	[37]	M. Mier, R. Ibañez, I. Ortiz, Influence of ion concentration on the kinetics of
13		electrodialysis with bipolar membranes, Sep. Purif. Technol. 59 (2008) 197-205.
14	[38]	Y. Wang, A. Wang, X. Zhang, T. Xu, Simulation of Electrodialysis with Bipolar
15		Membranes: Estimation of Process Performance and Energy Consumption, Ind.
16		Eng. Chem. Res. 50 (2011) 13911–13921.
17		