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Effectiveness of starch on microalgal biomass recovery, settleability and biogas production

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1 **Effectiveness of starch on microalgal biomass recovery,**
2
3 **settleability and biogas production**

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9 **Abstract**

10 In the context of wastewater treatment with microalgal cultures, coagulation-
11 flocculation followed by sedimentation is one of the suitable options for microalgal
12 harvesting. This process is enabled by the addition of chemicals (e.g. iron). However, in
13 a biorefinery perspective, it is important to avoid possible contamination of downstream
14 products caused by chemicals addition. The aim of this study was to evaluate the effect
15 of potato starch as flocculant for microalgal biomass coagulation-flocculation and
16 sedimentation. The optimal flocculant dose (25mg/L) was determined with jar tests.
17 Such a concentration led to more than 95% biomass recovery (turbidity<9NTU). The
18 settleability of flocs was studied using an elutriation apparatus measuring settling
19 velocities distribution. This test underlined the positive effect of starch on the biomass
20 settling velocity, increasing to >70% the percentage of particles with settling velocities
21 >6.5 m/h. Finally, biochemical methane potential tests showed that starch
22 biodegradation increased the biogas production from harvested biomass.

23 **Keywords** Coagulant, Flocculant, Harvesting, High rate algal pond, Microalgae, Starch.

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25 **1. Introduction**

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6 26 During the last decade the potential of microalgae as biorefinery feedstock has been
7
8 27 widely investigated (Uggetti et al., 2014). In spite of the promising results obtained,
9
10 28 scaling-up the technology is hampered by the high costs of the process. In particular, the
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12 29 biomass harvesting step represents 20-30% of microalgal biomass production costs
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15 30 (Barros et al., 2015).

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18 31 A number of solids separation techniques are currently available in the field of water
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20 32 treatment technology including centrifugation, flocculation and flotation (Uduman et
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22 33 al., 2010; Kurniawati et al., 2014) or membrane procedures such as magnetic, vibrating
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25 34 and rotating membranes. In general, for the production of low-value products such as
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27 35 biofuels, harvesting techniques should consist in low-cost and low-energy demand
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29 36 methods capable of processing a large volume of culture medium. Thus, coagulation-
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31 37 flocculation followed by sedimentation is among the most suitable options. This process
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34 38 is enhanced by the addition of chemicals such as salts of aluminum or iron. In a
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37 39 biorefinery context, it is important to ensure that downstream products are not
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39 40 contaminated by chemicals (Zheng et al, 2012). For this reason, the use of natural
41
42 41 organic flocculants like tannin based polymers or modified starch are being increasingly
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44 42 investigated (Vandamme, 2013). Indeed, potato starch could be seen as a residue from
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47 43 the potato industry (e.g. starch contained in potatoes peel).

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49 44 The aim of the present study was to evaluate the effect of potato starch on the
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51 45 coagulation-flocculation and sedimentation of microalgal biomass grown in a pilot high
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53 46 rate algal pond (HRAP) used for wastewater treatment. The optimal dose was
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56 47 determined with jar tests and the settleability of formed flocs was studied using an

48 elutriation apparatus measuring the settling velocities distribution. Moreover, the effect
49 of starch on biogas production was determined in biochemical methane potential (BMP)
50 tests.

51 **2. Materials and methods**

52 *2.1 Microalgal culture*

53 Microalgal biomass used for this experiment was cultivated in a pilot plant located at
54 the laboratory of the GEMMA research group (Universitat Politècnica de Catalunya,
55 Barcelona, Spain). The pilot plant consists of a HRAP in the form of a raceway pond
56 (0.5 m³ volume) fed with a continuous flow of 60 L/day of primary treated wastewater
57 (24 g chemical oxygen demand (COD)/ m²·d; 4 g ammonium nitrogen (NH₄⁺-N)/ m²·d).
58 The system had been in operation for 4 years prior to the experiment, and was described
59 in detail by Passos et al. (2013). At the time the experiments were conducted
60 microalgal populations were mainly composed by *Chlorella* sp. with a total suspended
61 solids (TSS) concentration about 200 mg/L.

62 *2.2 Flocculant*

63 Starch is a natural product having strong flocculating properties. Potato starch
64 solution 1% (C₆H₁₀O₅) provided by Panreac (Spain) was used as flocculant. Starch
65 addition did not modify significantly the pH of the system, which remained almost
66 constant along the experiments (9.5±0.6). The zeta potential was determined for the
67 selected starch concentrations and resulted in values of -35.8 mV (for 10 mg/L) and -
68 19.4 mV (for 25 mg/L).

69 *2.3 Jar tests*

70 The optimal dose of flocculant was determined by means of jar tests performed
71 following standard protocols (Metcalf and Eddy, 2003). Five jar tests were carried out
72 during February and March 2014, with starch concentrations from 5 to 80 mg/L.

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73 Turbidity and pH (measured with a HI93703 Hanna Instruments Turbidimeter and a
74 Crison 506 pH-meter, respectively) were determined from fresh HRAP mixed liquor at
75 the beginning of the experiments and from the supernatant liquid after the jar test.
76 Then, biomass recovery (RE) was calculated based on initial (Ti) and final (Tf) turbidity
77 measurements (Eq. 1). Recovery values of the five jar tests were then averaged.

$$78 \quad RE (\%) = (T_i - T_f) / T_i * 100 \quad (1)$$

79 *2.4 Elutriation test*

80 Elutriation is a water-current separation technique in which particles are washed out
81 according to their weight, volume or form. This test can be used to assess the feasibility
82 of separation treatment by settling (Krishnappan et al., 2004).

83 The water elutriation apparatus used in the present study was a modified version of a
84 system proposed by Walling and Woodward (1993). The system (Figure 1) consisted of
85 3 cylindrical settling columns of different diameter (corresponding to different settling
86 velocities) interconnected in series by glass and PVC tubing. The diameters of the
87 columns were: 50 mm in the first column, 100 mm in the second and 200 mm in the
88 third one. The sample entered the columns near the bottom and exited near the top,
89 allowing the sediment flocs that had settling velocities higher than the upward
90 suspension velocity to settle in the respective columns. Considering a flow rate of 0.21
91 L/min, the upward velocities generated in the three settling columns were 6.5 m/h, 1.6
92 m/h and 0.4 m/h, respectively. Thus, the first column collected biomass with settling
93 velocities >6.5 m/h, second column collected biomass with settling velocities between
94 6.5 and 1.6 m/h, and the third column collected the biomass with settling velocities
95 between 1.6 and 0.4 m/h. The outlet suspension contained the biomass fraction whose
96 settling velocity was lower than the suspension velocity in the third column (0.4 m/h).
97 All fractions of biomass retained in columns were collected and TSS analyzed

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98 according to Standard Methods (APHA AWWA-WPCF, 2001) for calculating the
99 settling velocity distribution.

100 Samples collected from the mixed liquor HRAP were mixed with starch at 2 different
101 concentrations (10 and 25 mg/L) and successively pumped through the series of settling
102 columns by means of a peristaltic pump. A sample without starch was also tested as
103 control. All tests were carried out in triplicate. In order to establish differences between
104 samples, ANOVA test was performed with Excel.

105 *2.5 Biochemical methane potential tests*

106 Biochemical methane potential tests were used to compare microalgal anaerobic
107 biodegradability and biogas production with and without flocculant addition. The BMP
108 test was performed in triplicate in samples with and without flocculant addition.
109 Digested sludge from a full-scale anaerobic reactor located in a municipal wastewater
110 treatment plant was used as inoculum.

111 BMP tests were performed in 160 mL bottles, in which 60 mg of microalgal biomass
112 (40.4 g VS/L) and 40 mg of digested sludge (30.4 g VS/L) were added, along with 2
113 different starch concentrations (10 and 25 mg/L). Three controls were also performed,
114 namely microalgal biomass and the 2 selected starch doses. All bottles were flushed
115 with Helium gas, sealed with butyl rubber stoppers and incubated at 35 °C until biogas
116 production ceased. Accumulated volumetric biogas production (mL) was calculated
117 from the pressure increase (periodically measured with a Greisinger GMH 3151
118 manometer), expressed under standard conditions. The net values of biogas production
119 and yield were obtained by subtracting the endogenous production of blank trials,
120 containing only inoculum.

121 **3. Results and discussion**

122 *3.1 Dose of flocculant*

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1
2 124 In the jar test, all tested starch doses led to high biomass recovery (RE>90%) (Figure
3
4 125 2). The optimal dose was 25 mg/L, which reduced turbidity from 151±14 to 6±2 NTU
5
6 126 (biomass recovery of 95.7%). On the other hand, increasing the starch concentration
7
8 127 further decreased biomass recovery (to 90.2% for a dose of 80 mg/L). The zeta potential
9
10 128 of 25 mg/L of starch (-19.4 mV) indicated an incipient instability of the solution (from
11
12 129 ±10 to ±30), with reduced electrical repulsion between particles. Jar test results indicate
13
14 130 that high biomass recoveries (about 95%) were also attained with doses lower than the
15
16 131 optimal. From a pragmatic point of view these doses would be preferred to avoid
17
18 132 overcosts. Even if the zeta potential of 10 mg/L (-35.8 mV) indicated only moderate
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20 133 instability of the solution with still strong electrical repulsion between particles, the
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22 134 dose of 10 mg/L (94% of recovery and 9 NTU) was also selected for settling velocity
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24 135 distribution and biochemical methane potential experiments.

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26 136 Considering that the initial biomass concentration was approximately 200 mg/L,
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28 137 from 0.05 to 0.15 g of starch per g of biomass were required to harvest more than 95%
29
30 138 of the biomass. Table 1 reports optimal doses for different chemicals, note that results
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32 139 comparison among different studies has to be taken with caution since the efficiency of
33
34 140 flocculants is dependent on the type of microalgae, their concentration and the medium
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36 141 conditions (Garde et al. 2014; Divakaran and Pillai, 2002).

37
38 142 According to Vandamme et al. (2010), increasing the initial biomass concentration
39
40 143 from 75 to 300 mg/L, increased the cationic starch required from 5 to 7.5 g per g of
41
42 144 biomass. Values higher than in this study corresponded to lower recovery efficiencies
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44 145 (85% vs. 95%). Higher recovery efficiencies (>95%) were obtained with cationic starch
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46 146 for the flocculation of *Scenedesmus dimorphus* in growth phases at the dosage of 10
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147 mg/L (Hansel et al., 2014). In that study, the initial culture turbidity was 50 NTU, which
148 corresponded to a biomass density of 0.12 g/L, lower than in this case (0.2 g/L).

149 When comparing doses it is important to consider the flocculants cost which can be
150 significantly different depending on the product. For example, chitosan (10US\$) is
151 about 10 times more expensive than cationic starch (1-3 US\$) (Vandamme et al., 2010).

152 *3.2 Settling velocity distributions*

153 The elutriation test was performed in order to determine the starch effect on the
154 settling velocity distribution of microalgal biomass. Results shown in Figure 3 represent
155 the percentage of biomass (expressed as TSS) retained in each column, which
156 corresponds to different settling velocities. Considering the control, which is microalgal
157 biomass without starch addition, 46% of the TSS were retained in the first column
158 (settling velocities >6.5 m/h), while 37% of the TSS were retained in the second column
159 (settling velocities between 1.6 and 6.5 m/h).

160 The effect of coagulation-flocculation on biomass settling velocities was very clear
161 since TSS retention in the first column increased from 46 % (in the control) to 78 and
162 73% for starch concentrations of 10mg/L and 25mg/L, respectively. This means that
163 starch raised settling velocities of biomass thanks to flocs formation. Results obtained
164 with 10 and 25 mg/L of starch were not significantly different.

165 Settling velocities have a direct impact on the settler dimensioning. In fact,
166 considering the results of this study, the addition of starch with the consequent increase
167 in settling velocity, would reduce the settler volume by 4 times maintaining the biomass
168 recovery efficiency $> 70\%$.

169 *3.3 Biochemical methane potential*

170 Considering anaerobic digestion as downstream process, the flocculant should not
171 have any inhibitory effect on biogas production. On the contrary, an organic

172 biodegradable flocculant such as starch may even increase biogas production. For this
173 reason, biochemical methane potential tests were carried out to compare the biogas
174 production obtained from control samples (microalgal biomass, and starch at 10 and 25
175 mg/L) and from samples with microalgal biomass and selected starch concentrations (10
176 and 25 mg/L). The results obtained (Figure 4) indicate that the addition of starch
177 increased the biogas production with respect to control microalgae biomass (503-536
178 vs. 467 mL biogas), even though the results are not significantly different ($p < 0.05$). The
179 comparison of microalgal biomass and starch codigestion (503-536 mL biogas) with the
180 digestion of both substrates separately (467 mL biogas for microalgae, 10-11 mL biogas
181 for starch) suggests certain synergic effect of the mixture which may be attributed to a
182 more balanced C/N concentration, enhancing to some extent the biogas production.

183 4. Conclusions

184 The effectiveness of starch as flocculant was tested on microalgal biomass. Low
185 starch doses (5-60 mg/L) allowed high biomass recovery $>92\%$. This is a promising
186 result, improving both microalgal biomass recovery and treated wastewater discharge.
187 The most appropriate starch dose was 25mg/L, leading to $>95\%$ biomass recovery and
188 an effluent turbidity <9 NTU. The elutriation test underlined the positive effect of starch
189 addition on the biomass settling velocity, increasing to $>70\%$ the percentage of particles
190 with settling velocities >6.5 m/h. Finally, biochemical methane potential tests show that
191 starch biodegradation increased the biogas production from harvested biomass.

192 5. Acknowledgements

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194 Competitiveness for financial support (DIPROBIO, CTM2012-37860). Raquel
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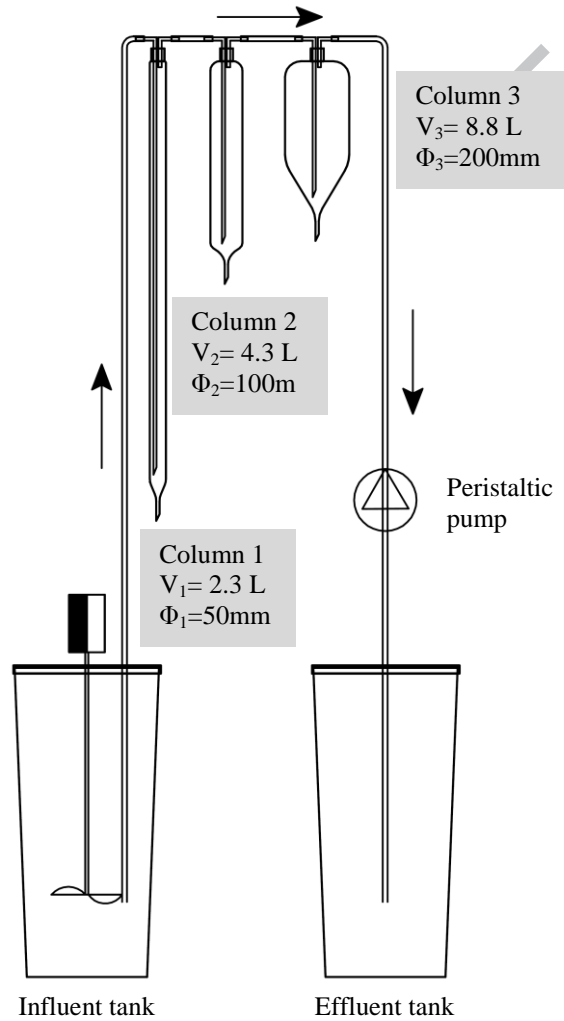
245 Tables and figures

Table 1. Literature results on microalgal biomass recovery using different chemicals.

Microalgal species	Microalgal concentration (g/L)	Chemical	Optimal dose (mg/L)	Biomass recovery (%)	pH	Reference
<i>Schizochytrium limacinum</i>	0.09	Aluminum sulfate	200	-		Gerde et al. 2014
	0.93			90		
	4.65			90		
	0.09	Cationic starch	10	37		
	0.93			80		
	4.65			80		
<i>Chlamydomonas reinhardtii</i>	0.03	Aluminum sulfate	250	90		
	0.31			90		
	1.06			60		
	0.03	Cationic starch	20	-		
	0.31			90		
	1.06			20		
<i>Scenedesmus</i> sp.	0.09	Aluminum sulfate	200	90		
	0.93			90		
	4.65			90		
	0.09	Cationic starch	40	70		
	0.93			90		
	4.65			90		
<i>Chlorella protothecoides</i>	0.44	Cationic starch	40	87	4	Letelier-Gordo et al. 2014
	0.56			95	7.7	
	0.77			96	10	
<i>Scenedesmus dimorphus</i>		Cationic starch	10-100	70-95		Hansel et al., (2014)
<i>Parachlorella</i> sp.	0.075	Cationic starch	20	90	5-	Vandamme et al., 2010
	0.15			90	10	
	0.3			85		
Mixture of						
<i>Spirulina</i> sp.,	10 NTU	Chitosan	15	60	4-9	Divakanar and Pillai, 2002
<i>Oscillatoria</i> sp.,	20 NTU			78		
<i>Chlorella</i> sp.,	30 NTU			82		
<i>Synechocystis</i> sp.	55 NTU			80		
<i>Nannochloropsis</i> sp.	665 × 10 ⁶ cell/mL	Chitosan	60	70-98	7-9	Farid et al., 2013
Mixture dominated by <i>Chlorella</i> sp.	0.2	Starch	25	95	8.9	This study
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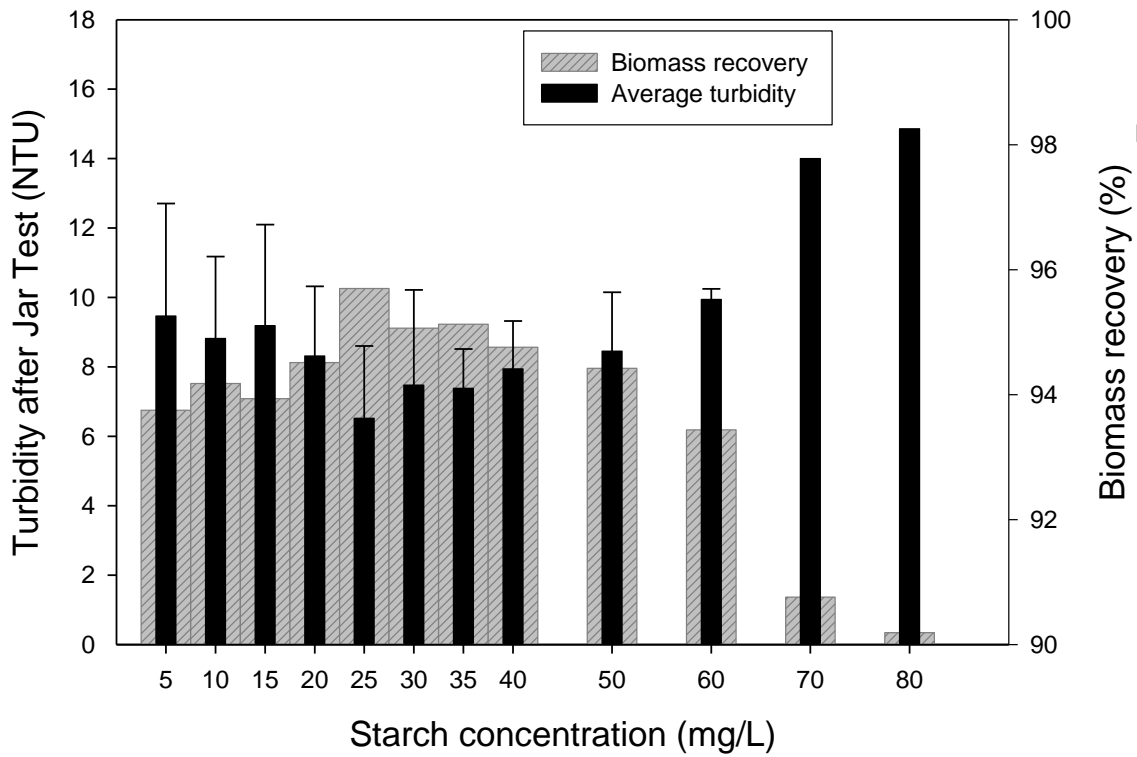
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251 Figure 1. Image and schematic view of the experimental elutriation apparatus used for

252 testing the settleability of microalgal biomass.

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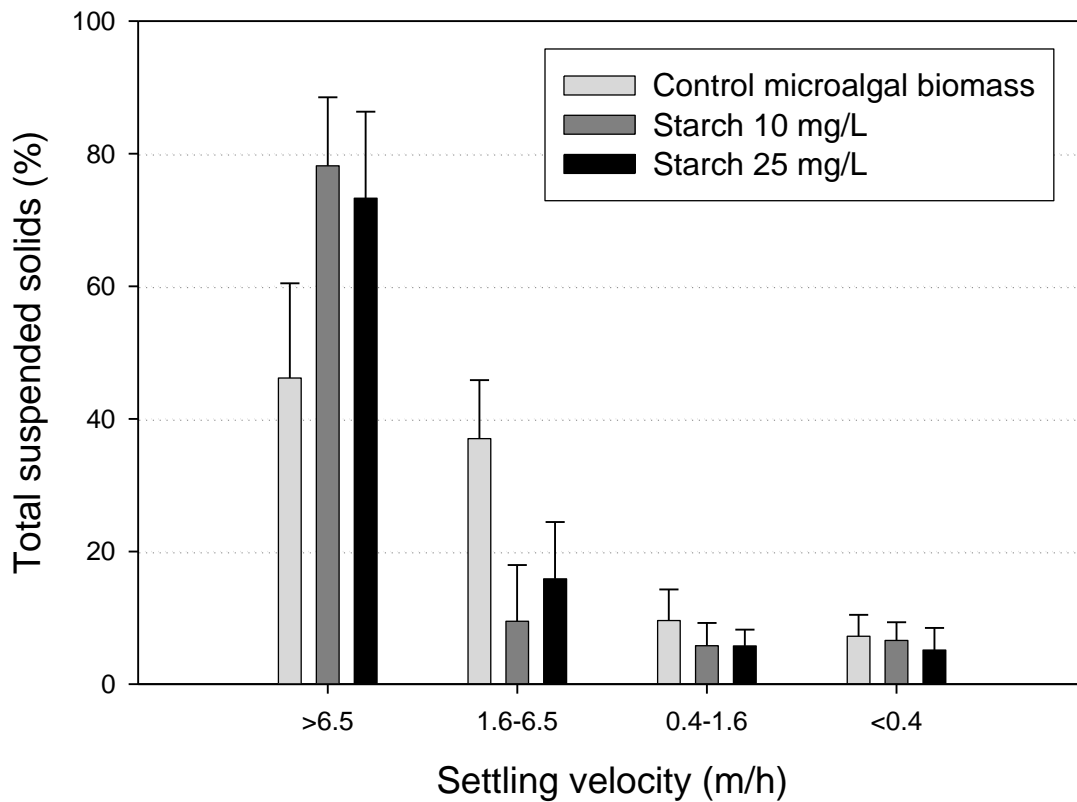
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255 Figure 2. Jar test results expressed as final turbidity and biomass recovery. Biomass

256 recovery was calculated from the initial turbidity (151 ± 14) and final turbidities shown

257 in the graph.

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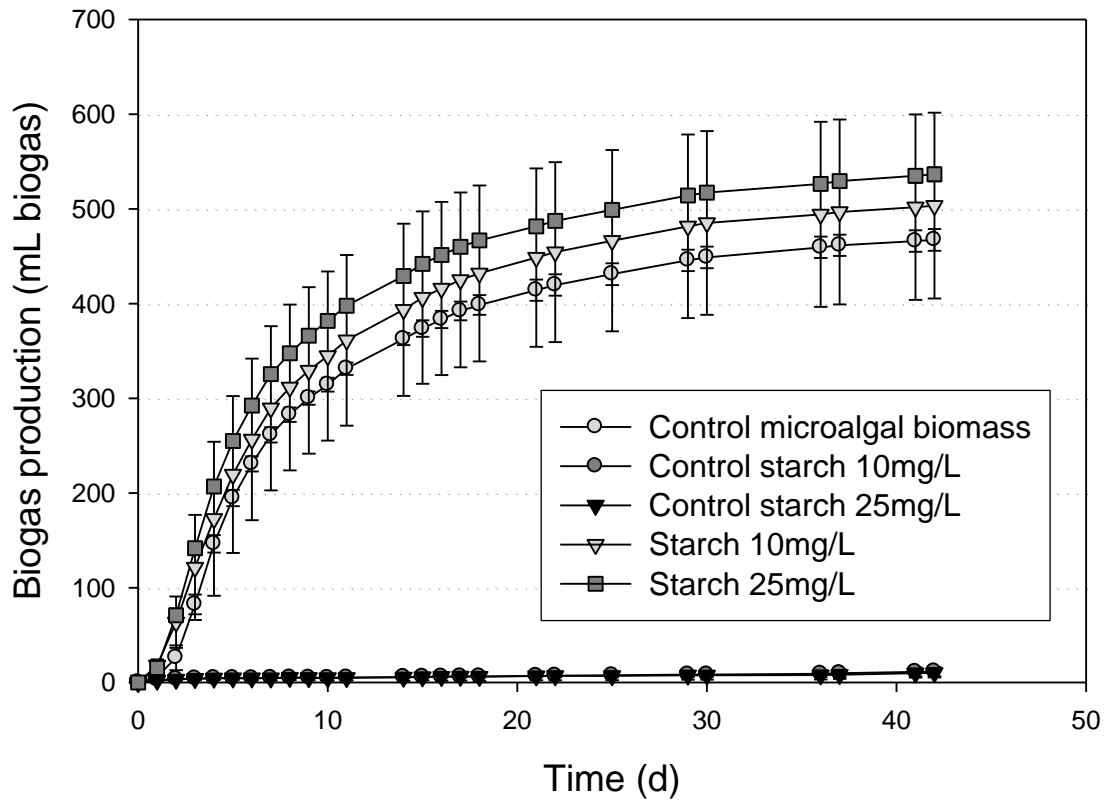


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260 Figure 3. Elutriation test results expressed as percentage of total suspended solids

261 corresponding to different settling velocities.

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264 Figure 4. Biochemical methane potential for control (microalgal biomass, starch 10
 265 mg/L and 25 mg/L) and microalgal biomass flocculated with starch doses of 10 and 25
 266 mg/L.

This study aims at evaluating starch as flocculant for microalgal harvesting

The optimal flocculants dose (25 mg/L) leads to more than 95% biomass recovery

Settleability was studied in elutriation apparatus measuring velocities distribution

Flocculants increased by 30% the particles with settling velocity higher than 6.5 m/h

Biochemical methane potential tests results show a biogas yield increment of 8-15%

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Figure 1. Image of the experimental settling columns used for testing the settleability of microalgal biomass.